Text S1. Supplementary methods, hydrocarbon, diatom and inferred primary production results and discussion. 1 2

3

4 **Supplementary Methods**

5 **Analyses of water chemistry:**

6 7 8 9 10 11 12 13 14 15 16 17 18 19 20 Water samples were collected from the central portion of each lake approximately 1 m below the surface, and were kept cool until transferred to the Taiga Environmental Laboratory (Aboriginal Affairs and Northern Development Canada) in Yellowknife, NT. The Taiga laboratory is an ISO/IEC 17025 accredited laboratory and a member of the Canadian Association for Laboratory Accreditation (CALA). Accreditation results can be viewed by entering member number 2635 on the CALA website. The Taiga Environmental Laboratory also regularly participates in QA/QC activities and exchanges with other accredited laboratories, including the National Laboratory for Environmental Testing (NLET; Environment Canada; Burlington, ON). Chemical determination followed a variety of standard protocols depending on specific analyte. Major ions were measured using ion chromatography following modified method SM4110B. Metals were analyzed by inductively coupled plasma – mass spectrometry following EPA method 200.8. Turbidity was analyzed by nephelometry following modified method SM2130B. True and apparent colour were analyzed via spectrophotometer following method SM2120C. Only chemical variables where the majority of samples occurred above the detection limit were included in further analyses.

21

22 **Sediment collection, radiometric dating and sediment age determination:**

 Sediment cores from the three study lakes were collected from a raft in the central, deepest portion of the lake using a Glew-type gravity corer (internal diameter: 7.62 cm; [1]) and sectioned at 0.5 cm resolution by vertical extrusion [2]. Sediment core intervals were dated using ²¹⁰Pb and ¹³⁷Cs radiometric techniques, via gamma spectroscopy [3, 4]. Selected sediment intervals were lyophilized, homogenized, entombed in epoxy resin and allowed to equilibrate for a minimum of two weeks. Samples were then analyzed for ²¹⁰Pb, ²¹⁴Bi and ¹³⁷Cs activity using an Ortec® well-type Ge crystal detector for 80,000 seconds each. Sediment age determination was conducted using the constant rate of supply (CRS) model [5], with the supported ^{210}Pb activity determined via the method established by [6]. The radioisotopic activities, as well as respective CRS-derived sediment ages, are presented in Fig. S1. For each core the relationship between the cumulative dry mass of the sediment interval and the unsupported ^{210}Pb activity (log transformed) was examined and found to be strong $(r > 0.8)$. 23 24 25 26 27 28 29 30 31 32 33 34

35

36 **Supplementary PAH methods:**

37 38 39 40 41 42 43 44 45 Wet sediments were homogenized and mixed in an approximate 1:1 ratio with Agilent brand Hydromatrix[®]. The samples were spiked with ¹³C labelled PAHs (Cambridge Isotope Laboratories). PAHs were extracted from the wet sediments using accelerated solvent extraction (ASE, Dionex) at 100˚C with hexane:dichloromethane (DCM), followed by 35% acetone:65% hexane. The non-polar extract (PAHs, hexane and DCM) was separated from the polar extract (water and acetone) by combining the entire sample with DCM, 3% NaCl and saturated Na₂SO₄ in a series of liquid-liquid extractions. Sulphur compounds and pigments were removed from the final extract on an Agilent 1100 Preparative Liquid Chromatograph using Waters Envirogel Columns and USEPA method 3640A. The samples were evaporated to 1ml and fractioned,

for 244 for C3-fluroanthene and C3-pyrene, m/z 258 for C4-fluoranthene and C4-pyrene, m/z 242 for C1-benz(a)anthracene and C1-chrysene, m/z 256 for C2-benz(a)anthracene and C2 chrysene, m/z C3-benz(a)anthracene and C3-chrysene, m/z 266 for C1-benzfluoranthene and C1 benzopyrene , m/z 280 for C2-benzofluorantene and C2-benzopyrene. Quantification was performed using 13 C labelled non-alkylated PAHs (Cambridge Isotope Laboratories) for all of the compounds listed above. 69 70 71 72 73 74

75

76 **Supplementary Results and Discussion**

77 **Present-day lakewater chemical comparisons:**

78 79 80 81 82 83 84 85 86 87 88 89 90 91 Comparisons of present-day physical and chemical limnological variables reveal major differences between the three *a priori* defined groups (all comparisons made using two-way ANOVA, df=2, 98, followed by Tukey HSD post-hoc test run on normalized data) (Fig. S2). Drilling sump lakes exhibited significantly higher Cl concentrations than either permafrost thaw or control lakes (F=7.91, p<0.001). Permafrost thaw lakes exhibited significantly increased K^+ $(F=23.6, p<0.001)$, Ca^{2+} (F=50.2, p<0.001), and Na⁺ (F=30.7, p<0.001) concentrations, specific conductivity (F=47.3, p<0.001) and total dissolved solids (F=18.8, p<0.001) than the drilling sump lakes which were all significantly greater than in the control lakes (Fig. S2). The drilling sump lakes were significantly shallower than the permafrost thaw or control lakes $(F=19.7,$ $p<0.001$). Lakes with permafrost thaw slumps had significantly higher apparent colour (F=37.0, p<0.001), SO_4^2 concentration (F=36.0, p<0.001) and lower total suspended solids (F=4.52, p=0.01) than the other two groups. The control lakes had lower pH than either the drilling sump or permafrost thaw lakes $(F=22.1, p<0.001)$. No significant difference was observed in water turbidity between the three groups $(F=1.46, p=0.24)$. Analysis of similarity (ANOSIM) and

similarity percentages (SIMPER) were used in order to assess differences in the *a priori* defined groups, as well as determine which variables contributed to any statistical differences. SIMPER analysis, run on a Euclidean distance matrix, determined that Cl- , Si, and Mn concentrations, as well as maximum depth, contributed significantly to the difference between the drilling sump and control lakes as well as the drilling sump and permafrost thaw lakes (Table S2). Ca^{2+} , Si, and Mn concentrations, as well as specific conductivity, were found to contribute to the dissimilarity between the permafrost thaw and control lakes, though notably chloride concentration was not observed to be important (Table S2). 92 93 94 95 96 97 98 99

100

101 **Polycyclic aromatic hydrocarbons (PAHs):**

102 103 104 105 106 107 108 109 110 111 112 113 Sedimentary concentrations of the 12 priority PAHs, including the most common PAHs from petrogenic sources, did not change at the time of, or subsequent to, drilling sump construction in any of the three study lakes (Fig. S3). Included in this summation are the following compounds: phenanthrene, anthracene, fluoranthene, pyrene, benz[a]anthracene, chrysene, benzo[b]fluoranthene, benzo[k]fluoranthene, benzo[a]pyrene, indeno[1,2,3-cd]pyrene, dibenz[a,h]anthracene, and benzo[g,h,i]perylene. This suggests that, unlike the impact of largescale extraction operations such as the Alberta oilsands [7, 8], impacts from exploratory drilling activities in this region of the Arctic do not appear related to contamination by hydrocarbons themselves. This is not surprising, given that these exploratory operations extract comparatively few hydrocarbons during the process of test well development, and given that relatively high, stable background PAH concentrations were observed in the sediments of all three lakes throughout the past few hundred years (Fig. S3).

114

115 **Sedimentary diatoms:**

116 117 118 119 120 121 122 123 124 125 126 127 128 129 130 131 132 133 134 135 136 137 Sedimentary diatom assemblages from Lake I20, impacted by a failing drilling sump, do not show any response coincident with, or subsequent to, sump construction (Fig. S4). The modern assemblage is dominated by the planktonic taxon *Cyclostephanos tholiformis*, as well as a variety of small benthic taxa of the group *Fragilaria sensu lato* (primarily *Staurosirella pinnata*, *Staurosira venter* and *Pseudostaurosira brevistriata*). Prior to approximately 1950, when a gradual increase in *C. tholiformis* began, the assemblage was dominated by small benthic fragilarioids at greater than 70% relative abundance (Fig. S4). Similarly, in Lake C23, the assemblage is dominated by small benthic fragilarioid taxa throughout the recent past. At ~1910 an increase in planktonic taxa, including *Discostella pseudostelligera* (synonymous with *Cyclotella pseudostelligera*), as well as a number of pennate planktonic taxa occurred, coincident with the onset of warming in this region as recorded by several independent proxies [9]. The primary diatom changes in both I20 and control Lake C23 occur earlier than the known construction of drilling sumps in the vicinity of the lakes, and are characterized by increased planktonic diatom abundance, indicative of climate warming [10], similar to the timing and nature observed in many other lakes in this region [9]. The changes in diatom assemblages in I20 and C23 occur simultaneously with increases in inferred primary production (Fig. S3), suggesting an influence of warming on overall primary production in these lakes. No major changes were observed in the diatom record from C1A (Fig. S4), which is composed of an assemblage dominated by small *Fragilaria sensu lato* taxa as well as a variety of other periphytic and benthic species. The inferences of primary production in C1A were below detection limits (Fig. S3). In no lake was an increase in saline-tolerant taxa observed, such as those recorded following saline intrusion in the nearby outer Mackenzie Delta [11]. The lack of diatom response

178 palaeolimnological analysis. The Holocene 22:1447-1456.

 Σ 12 priority PAH concentration (ng PAH/g OC)

A Sump-impacted Lake I20

_B Control Lake C23 C Control Lake C1A udels Grade Of Think Into a real

lanktonic tata

 $\overline{2}$ 20 40 60.80

Benthic tradition of late

	Sur Area*	Catch Area*	$Z_{\rm max}$ *	Alk*	$Col*$	$Cond*$	pH	TDS*	Turb*	Ca	Mg	Na	K	Cl	SO_4	Fe	Mn
	(ha)	(ha)	(m)	(mg/l)	CU)	$(\mu S/cm)$		(mg/l)	(NTU)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	$(\mu g/l)$	$(\mu g/l)$
Drilling-sump lakes $(n = 20)$																	
Mean	15.8	83.5	1.7	58	81	214	7.8	160	2.5	23.2	10	8.1	1.8	22.8	8	477	46
Median	4.8	27.3	1.5	49	68	144	7.8	132	1.8	18.9	7	4.9	1.3	8.0	6	394	31
STD	22.4	162.9	0.8	35	51	155	0.5	110	2.1	13.2	6	8.1	1.6	42.9	9	397	50
Max	85.7	749.4	3.4	145	243	738	9.0	554	9.6	54.8	27	37.3	6.7	198.0	40	1570	214
Min	0.5	15.5	1.0	12	19	58	6.9	61	0.5	6.9	33	1.8	0.5	1.3		50	9
Thaw slump-affected lakes $(n = 34)$																	
Mean	13.1	50.6	5.7	90	20	404	8.1	275	2.8	49.9	17	11.6	2.5	9.1	102	186	36
Median	6.2	26.0	5.0	86	16	354	8.1	192	1.5	42.3	13	10.9	2.1	6.7	52	131	21
STD	22.6	62.6	3.1	41	14	243	0.3	251	3.1	34.2	12	6.8	1.3	5.9	120	190	47
Max	116.5	254.4	14.2	174	50	1000	8.4	1330	15.8	146.0	53	30.8	6.1	20.6	406	855	253
Min	1.2	4.7	1.7	20	5.	84	7.2	52	0.5	6.9	3	2.3	0.4	1.9	2	50	4
Control lakes $(n = 47)$																	
Mean	5.0	64.8	3.9	29	84	107	7.3	99	3.2	11.8	4	3.7	1.0	6.1	10	895	42
Median	2.9	23.0	3.3	26	74	102	7.5	84	2.5	11.3	4	3.3	1.0	5.9	4	505	25
STD	6.3	151.9	2.8	18	62	54	0.6	93	2.5	5.8	2	1.8	0.5	3.9	19.	1058	71
Max	32.5	749.4	11.2	94	266	287	8.2	694	12.6	27.9	9	7.7	2.5	13.6	118	5960	482
Min	0.5	3.9	1.0	4	5.	29	4.3	40	0.3	4.2		0.8	0.2	0.9		$\overline{0}$	\overline{c}

Table S1: Summary statistics of water chemistry data. Sites were sampled in the summers of 2005 and 2007, with samples taken at approximately one metre water depth.

Table S2: Results of similarity percentages (SIMPER) analysis.